



Seasonality and Climatic Factors Affect Diversity and Distribution of Arthropods Around Wetlands

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ABSTRACT

The diversity and distribution of arthropods associated with seasons and environmental variables in two different wetlands sites was estimated and study was performed on monthly basis for a whole year. The arthropods were collected by sweep net, direct hand picking, forceps. For each sampling day in each month, 20 sweeps were randomly taken in each block. The insects at rest or on shrubs were manually collected. Overall, 5867 individuals were recorded pertaining to 152 species. Among the arthropods collected from two fields, number of the order Diptera was most diverse followed by those of Coleoptera, Hymenoptera, Hemiptera, Orthoptera, Lepidoptera, Araneae and Neuroptera. The number of arthropod species differed between localities and between the studied seasons. The arthropod composition varied significantly according to the sampled sites and according to the seasons studied. Temperature was positively correlated to diversity while humidity was negatively correlated to diversity and abundance. Results might be beneficial in developing the strategy for natural biological control. This can only be achieved by coordinated planning and conservation measures among sites.

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Authors' Contribution

WM and NR conceived and designed the experiments. WM and SN performed the experiments. WM, EBAK and SN analyzed the data. WM and EBAK wrote the manuscript.

Key words

Diversity, Environmental variables, Wetlands, Ecological indicators, Pakistan

INTRODUCTION

We all are well aware that climate change is occurring (Houghton *et al.*, 2001), and these may affect the diversity and distribution of various organisms (Bale *et al.*, 2002). Population dynamics is dominant challenge to ecology to drive population cycle, response of species to ecological variation can interact with density dependent feedback processes (Crozier, 2004; Zamani *et al.*, 2006; Savage *et al.*, 2004; Amarasekare and Sifuentes, 2012). Negative feedback mechanism is delayed due to the intricacies in life cycles (e.g. intraspecific competitions), which provokes the series in population cycles (Murdoch *et al.*, 2003; Amarasekare and Coutinho, 2014; Nisbet and Gurney, 1983; Nisbet, 1997).

Arthropods play consequential function in ecosystems service, including; predatory role, pollination processes, nutrient recycling, parasitoid behavior and in the sustainability of the ecosystem. Alteration in insect biodiversity and richness has threats of changing the services being provided by these tiny creatures (Hillstrom and Lindroth, 2008).

Diversity and distribution of fauna is directly and

indirectly affected by abiotic variation. It is noteworthy that, species are declining with regards to amend in environmental factors. Reproduction, development, interaction and historical traits (Dannon *et al.*, 2010; Englund *et al.*, 2011; Lang *et al.*, 2011) and conservation of species is mainly impacted by direct effect of abiotic factors (Hou and Weng, 2010; Jandricic *et al.*, 2010; Urbaneja *et al.*, 2001; Castillo *et al.*, 2006; De Conti *et al.*, 2010; Nishikawa *et al.*, 2010; Morgan *et al.*, 2001; Bommireddy *et al.*, 2004; Medeiros *et al.*, 2003; Ragland and Kingsolver, 2008; Parajulee, 2006; Matadha *et al.*, 2004; Huang *et al.*, 2008; Ulmer *et al.*, 2006). Indirect effects of environmental variation also putting the values down of arthropods abundance and distribution. Considerate as the atypical environment, meaning i.e. climate change, influence the density of a specific population by exerting direct and indirect effects (Bale *et al.*, 2002; McMahon *et al.*, 2011; Deutsch *et al.*, 2008; Sheldon *et al.*, 2013; Root *et al.*, 2003; Kingsolver *et al.*, 2011). Present study site has been chosen to expound the mechanism that drive species, because seasonal and climatic variation are operating the ecosystem as in changing the sequences of fauna and less known diversity and abundance of arthropods.

Climatic variations are giant dilemma in different ecosystems; as seasonal extremes alter the diversity and abundance of arthropods in any territory. The common

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aspects that affect species diversity in any ecosystem include habitat structure climate and biogeographical dynamics (Dossey, 2010). Wet territories have significant role in the strengthen of ecosystem, and in the amelioration of fauna. Arthropods play eminent role in pollination (Berenbaum *et al.*, 2006; Thompson and McLachlan, 2007), biological control agents of pest (Choate and Drummond, 2011). Arthropods are effective indicator of ecological variations due to their miscellaneous natural features and necessities (Hall and Castner, 2000; Clarke *et al.*, 2008; Jaganmohan *et al.*, 2012).

Alterations in seasons and climate have a great influence on insects and plants association at various aspects; either directly, by change in physiological aspects or indirectly by morphological changes endured by hosts plant (Morrison and Morecroft, 2006) along with the thruway of abundance, biodiversity, evenness and richness (Thuiller *et al.*, 2005). Diversity of the arthropods is influenced by the season rather than the age of the plantation as well as seasonality have a great impact on diversity, abundance, richness and evenness of arthropod fauna (Liu *et al.*, 2013). Seasons prevail different substantial distributional patterns of terrestrial arthropod community (Doblas-Miranda *et al.*, 2007). Some studies demonstrated the impact of shrubs age on temporal-spatial distribution of biodiversity of arthropods community at various micro-environment in different season (Whitford, 2000).

Recent regional reports and trends in biomonitoring suggest that insects are experiencing a multicontinental crisis that is apparent as reductions in abundance, diversity, and biomass (Forister *et al.*, 2019; Wagner, 2018; Dirzo *et al.*, 2014; Janzen and Hallwachs, 2019; Sánchez-Bayo and Wyckhuys, 2019). Further work needs to be done in the study area by expanding the scope and duration of the study area and also, by employing different sampling techniques. Motivated by a distinguishing pattern observed in the dynamics of population, we develop a hypothesis: Population dynamics/ diversity and distribution can be directly and indirectly influence by seasonality and abiotic factors. This study focuses on the diversity of distribution of arthropods population in accordance with seasonality and climatic variations.

MATERIALS AND METHODS

Sampling area

The research area was selected from two different water bodies of Faisalabad, Pakistan: Lake and Canal territories (Table I). Faisalabad is present at a height of 182.88 m above ocean level, having the latitude of 30° to 31.5° N and at longitude of 73° to 74° E (GOP, 2016).

Sampling and identification

The sampling was made from the selected fields for one year (four seasons, monthly from each territory between January to December 2018). The foliage fauna was collected from the fields (Lake and Canal) territory for 02 hours from 08:00-10:00 hr. The arthropods were collected by Sweep net, direct hand picking, forceps. For sampling day in each month, 20 sweeps were randomly taken in each block (Quadrates) while insects at rest or on shrubs and soil surface were manually collected.

After collection specimens were put in killing jars. Identification were made based on microscopes observations, following taxonomic keys (Borror and DeLong, 2005; Triplehorn and Johnson, 2005; Rafi *et al.*, 2005). Subsequently, recognized specimens were specifically organized in the tabular arrangement as specified by their taxonomy e.g. order, family, genus, species on the base of their morphological characteristics. During the whole trial temperature and humidity were also noted of the sampling day (these variables were noted before the start and at the end of sampling and were averaged/mean) for the correlation analysis with species diversity.

Statistical analysis

Diversity related various issues were calculated according to Shannon's Diversity Index (Shannon, 1948). Thereafter, all the observed specimens were arranged in table form according to their morphological and taxonomic characters. Variations in arthropod richness according to seasons and localities were evaluated using two-way ANOVA, followed by Tukey's pairwise multiple comparison tests. Variation in species composition was evaluated using a Permutational Multivariate Analysis of Variance (PERMANOVA; Anderson, 2001). The similarity distance was calculated by the Bray-Curtis index. Statistical significance was obtained through comparisons with a null model (4999 permutations of the original matrix). To illustrate these possible differences in species composition was conducted a non-metric multidimensional scaling (NMDS). These analyses were performed in R software v. 3.5.0 (R Development Core Team, 2018) using vegan 2.3-4 package (Oksanen *et al.*, 2015). We graphically analyzed the residuals for the assumptions of the tests performed, including the normality of errors and the homogeneity of variances. All tests were analyzed at the level of significance $\alpha=0.05$. Additionally, to sustain the reliability and to decrease uncertainty, data from the four seasons were also combined and used to present part of the results. The following metrics were used to estimate diversity: Shannon's diversity index, species richness, species dominance, species evenness and community dominance (Shannon, 1948; Magurran, 1988).

Table I. Characterization of the two different water bodies studied in Faisalabad, Pakistan.

Information	Canal	Lake
Coordinates	31.4504° N, 73.1350° E	31.4504° N, 73.1350° E
Extension	5 hectares	5 hectares
Vegetation	Agriculture crops, grassland, plantation (small and large)	Grassland, herbs and shrubs
Soil characteristics	Silt and sandy loam, pH 8-9, decaying organic matter 55-65 %	Silt and sandy loam, deep and dark, decaying organic matter 55-65 %
Distance from water bodies (Area in which sampling is done)	100-1000 m	100-1000 m
Distance between water bodies	12 Kilometers	

RESULTS

Overall, 5867 specimens were documented and resulted that abundance as well as species diversity was maximum for Lake 61.63% (n= 3616), while minimum was recorded for canal 38.37% (n= 2251). Whole data pertaining to 2 classes, 8 orders, 61 families, 91 genera, 114 species and 2 classes, 7 orders, 42 families, 61 genera, 77 species from lake and canal respectively (Supplementary Table I). From Lake, the highest abundance was recorded for *Tanytarsus* spp. 17.56% (n=649) followed by *Exitianus indicus* 6.66% (n = 241) and from canal maximum was recorded for *Musca domestica* 8.22% (n = 185) followed by *Lucilia sericata* 6.09% (n = 137), while certain species had less abundance or found only at single site. At the order level, Diptera remained extraordinary 46.47% and 37.39% from lake and canal respectively, followed by other orders in (Fig. 1). Variations in species abundance were observed between different seasons for both sites. It was noted that species showed significant variation among different seasons viz. From lake fields (spring, summer, autumn, winter) seasons abundance was noted 36.89%, 20.46%, 28.29%, 14.35% respectively. From canal fields, abundance was documented as for all seasons 34.87%, 20.12%, 29.41%, 15.59% respectively.

The temperature was significant and positively correlated with the diversity of arthropods (Lake territory: $r = 0.502$, $p < 0.05$; Canal territory: $r = 0.481$, $p < 0.05$) and humidity was significant negatively correlated with diversity (Lake territory: $r = -0.334$, $p < 0.01$; Canal territory: $r = 0.481$, $p < 0.01$) (Table II).

From Lake species abundance varies 519 (Winter) from 1334 (Spring), while in Canal abundance varies 351 (winter) from 785 (spring) (Table III). Diversity (H'): 3.62 was the maximum diversity recorded from the Lake territory, whereas, from Canal the maximum value was 3.55 in summer for both territories. Dominance (D): dominance was recorded maximum for spring season of Lake 0.088

and for Canal maximum 0.066 was documented in winter season. Evenness (J): Highest evenness was recorded highest 0.778 from Canal fields and maximum for Lake fields 0.616. Richness S: Richness was noted maximum 10.44 and minimum 7.651 for spring season of Lake and Canal respectively. Community dominance: in the spring season of both territories CD% was recorded maximum. Other diversity indices were also noted (Table III).

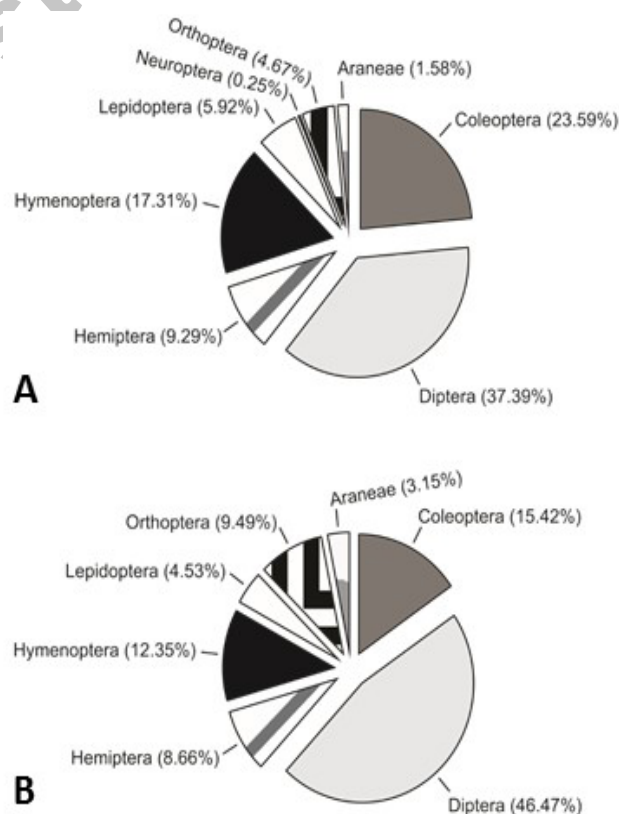


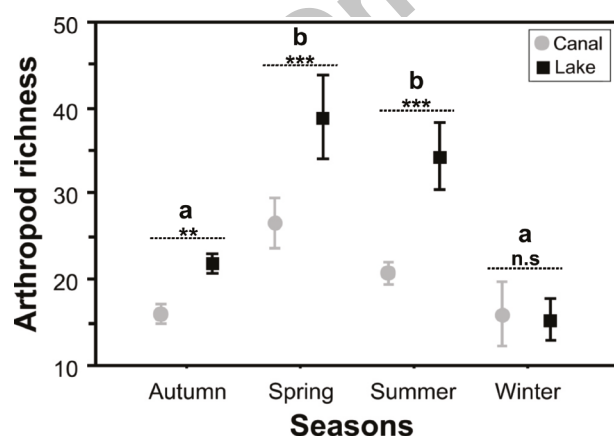
Fig. 1. Diversity and abundance of insect fauna in lake (A) and canal (B) territories.

Table II. Correlation analysis of species with abiotic factors (temperature, humidity). * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$, n.s= no significant difference.

Factors	Lake			Canal		
	Abundance	Temperature	Humidity	Abundance	Temperature	Humidity
Abundance	1	0.502*	-0.334 ^{NS}	1	0.481*	-0.607**
Temperature		1	-0.765**		1	-0.736**
Humidity			1			1

Table III. Diversity, abundance, dominance, evenness and richness of species recorded in the two water bodies (canal and lake) studied in Faisalabad, Punjab, Pakistan.

Diversity factors	Lake				Canal			
	Spring	Summer	Autumn	Winter	Spring	Summer	Autumn	Winter
Individuals	1334	740	1023	519	785	453	662	351
Taxa richness S	79	72	40	30	52	45	36	31
Dominance D	0.088	0.060	0.058	0.051	0.063	0.035	0.064	0.066
Simpson 1-D	0.911	0.939	0.9414	0.938	0.936	0.964	0.935	0.933
Community dominance%	338.94	159.08	162.38	61.29	128.30	46.18	107.78	69.69
Diversity H	3.3	3.62	3.206	3.007	3.234	3.556	3.097	3.042
Evenness $e^{H/S}$	0.343	0.518	0.616	0.674	0.488	0.778	0.614	0.675
Brillouin	3.189	3.445	3.12	2.894	3.11	3.369	2.987	2.881
Menhinick	2.16	2.647	1.251	1.317	1.856	2.114	1.399	1.655
Magarlef	10.84	10.75	5.6277	4.639	7.651	7.194	5.389	5.119
Equitability J	0.7552	0.846	0.869	0.884	0.818	0.934	0.864	0.885
Fisher_alpha	18.38	19.72	8.293	6.929	12.52	12.42	8.168	8.202
Berger-parker	0.2459	0.206	0.149	0.0847	0.151	0.068	0.145	0.170
Chao-1	79.75	72	40	30	52.1	45	36	31

**Fig. 2.** Arthropod richness observed according to seasons and water bodies (canal and lake) studied. Different letters above error bars represent significant differences between stations. * $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$; n.s, no significant difference.

The number of arthropod species differed between localities ($F_{1,16} = 13.937$, $p < 0.001$) and between the studied seasons ($F_{3,16} = 13.966$, $p < 0.001$). No significant interaction was found between these factors ($F_{3,16} = 0.1003$, $p < 0.497$). For both localities, no significant differences in species richness were observed between the Autumn and Winter seasons, as well as between Spring and Winter (Tukey Post Hoc-Test > 0.05). All other stations were significantly different from each other (Tukey Post Hoc-Test < 0.05 ; Fig. 2).

The arthropod composition varied significantly according to the sampled sites ($F_{1,22} = 3.9348$, $p < 0.001$) and according to the stations studied, for both Canal ($F_{3,8} = 1.763$, $p = 0.005$) and Lake ($F_{3,8} = 2.1702$, $p < 0.001$; Fig. 3). When considering the presence and absence of different arthropod species, the similarity observed between the fauna of the two localities was less than 20% (Fig. 4). The greatest similarity found between seasons was between summer and spring for both locations (about

50% similarity for Canal and 45% similarity for Lake; Fig. 4). Winter was being presented with the lowest similarity with the other (Fig. 4).

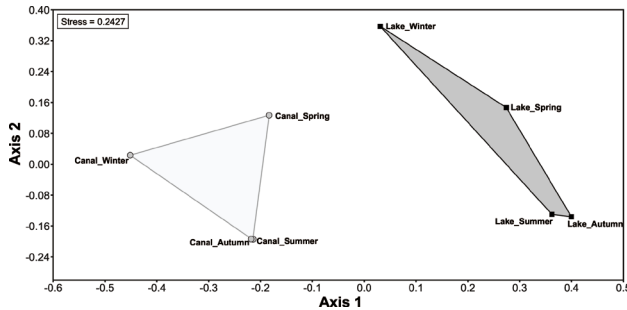


Fig. 3. Non-metric multidimensional scaling- NMDS (Bray-Curtis distance) of arthropod species composition according to different seasons and water bodies (canal and lake) studied. Stress = 0.2427

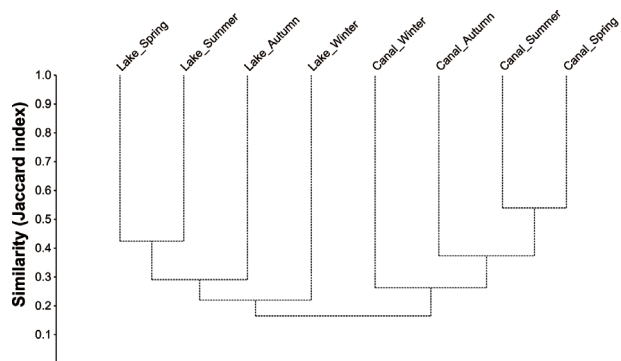


Fig. 4. Jaccard similarity dendrogram showing the similarity in arthropod fauna between the different seasons of the two water bodies (canal and lake) studied in Faisalabad, Punjab, Pakistan.

DISCUSSION

Biodiversity can be considered the whole variety of life, at all levels of the organization, regardless of the classification form (Colwell, 2009). Understanding this diversity has been the interest of several studies, both in aquatic and terrestrial environments (Footitt *et al.*, 2009). Climatic variations are giant dilemma in our ecosystem; as seasonal extremes alter the diversity and abundance of arthropods in any territory. The common aspects that affect species diversity in terrestrial and aquatic ecosystem include habitat structure climate and biogeographical dynamics such as habitat area (Dossey, 2010). Wet territories have significant role in the strengthen of ecosystem, as they play eminent role in pollination, biological control of the pest etc. (Thompson and McLachlan, 2007). Arthropods

are effective indicator of ecological variations because of their miscellaneous natural features and necessities (Hall and Castner, 2000; Clarke *et al.*, 2008; Jagannmohan *et al.*, 2012).

In this research, a total of 5867 specimens were recorded containing 8 orders, 72 families and 152 species (Table II). Similar study was done up to order, family and species by Majeed *et al.* (2019) and found that high population dynamics pertains alongside the water body. The role of arthropods fauna in the betterment vegetation alongside the wet territory needed to know due to the future conservation perspective of wetland fauna in biodiversity and ecology as well as in reverse vegetation also play important role in the enhancement of faunal diversity (Mahmoud and Shebl, 2014).

In our results, from Lake, maximum population was recorded for *Tanytarsus* spp. followed by *Exitianus indicus*. While for the Canal locality, maximum population was recorded for *Musca domestica* followed by *Lucilia sericata*. These species are in fact known to have high abundance of individuals (Basset, 2001). Romeny *et al.* (2002) documented that more diversity was recorded for *Chironomid* midges 51% while other species were present in less number. Similar to the finding in other studies (Maguire, 2006; Vincent and Ring, 2009; Michael *et al.*, 2007) diversity and abundance of species at order level was also examined and noted that Diptera had more influence on other recorded order (Romeny *et al.*, 2002). It was found that Diptera presents to be more when compared to other species. The supremacy of Coleoptera and Lepidoptera in our study areas is showed by the extraordinarily maximum abundance of Ptiliidae, which prefer rotting wood, moist soil, and litter, are a prospective bioindicator of moist habitats (Sawada and Hirowatari, 2002; Hall, 2001; Sorensson, 2003). Ghosh-Harihar (2013) stated the diversity, richness and abundance of foliage arthropods. This trend has also been documented by Veenakumari and Prashanth (2009) who studied insect diversity in the mangroves of Andaman and Nicobar Islands of India who reported 50% Lepidoptera; 5% Orthoptera; 3% Hymenoptera; 20% Coleoptera; 15% Hemiptera; 5% Diptera; 2% Thysanoptera.

Habitat structure is the amount, composition and arrangement of biotic and abiotic factors in any area, as an empirical driver of many ecological processes and regulate communities of many organisms in any ecosystem (Lovett *et al.*, 2005). Seasonal changes and its effect on diversity was also noted significantly *viz.* from lake fields (spring, summer, autumn, winter) seasons abundance was noted 36.89%, 20.46%, 28.29%, 14.35%, respectively. From canal fields, abundance was documented as for all seasons 34.87%, 20.12%, 29.41%, 15.59%, respectively. Maguire

(2006) studied the influence of abiotic factors on the species diversity and distribution in any habitat. Effects of temperature and resource variation on insect population dynamics was also observed by Johnson *et al.* (2016). In the present study a significant positive relationship was found between arthropod diversity and temperature. While humidity showed a significant negative relationship with diversity from lake and canal territory respectively. Arun and Vijayan (2004) estimated the relationship among ecological factors and the abundance of insects. Environmental factors such as rainfall, maximum temperature, minimum temperature, and relative humidity were monitored during the study. Among them, only the minimum temperature presented a significant negative correlation with the monthly variations in insect abundance.

The density of arthropods varied greatly over the growing season. Most arthropod groups were abundantly found in the spring-summer growing season (Rango, 2005). Liu *et al.* (2013) researched on arthropod distribution at microenvironments along vegetative areas in different seasons including shrubs of various plants and evaluated their activity periods as summer, spring and autumn. Tanaka and Tanaka (1982) estimated the changes in the arthropod's diversity due to rainfall and seasons and found that seasonal changes have significant correlation with the abundance of arthropod fauna.

In the present study, the lowest abundance values were found for winter. For species richness we found that, among seasons, spring has highest richness and evenness of species. Additionally, a marked variation in diversity indices was observed between the territories. These results are similar to previous studies that evaluated the diversity of arthropods in other localities. (Balakrishnan *et al.*, 2014; Tewari *et al.*, 2006). The diversity indices used in this study were complementary and showcased the high diversity and richness of insects in the ecosystem. Plant-insect interaction by way of phytophagy and mutualism may be responsible for this high species diversity as it has been shown that the diversity of vegetation in a habitat influences the diversity of insect species in such habitat (Alarape *et al.*, 2015).

CONCLUSIONS

The pattern of arthropod diversity was investigated around two different wetlands and concluded that maximum abundance was recorded for lake fields as well as species number remained extraordinary as compared to Canal. The species richness of order Diptera was positively influenced by degree of habitat complexity. Not only the abundance, but species diversity was also positively correlated to the foliage vegetation structure. Seasonal alteration in the

number of species and specimens was also noted during the whole trial. Hence, this study shows that seasons, climatic conditions, habitat structure have a significant impact on. Results might be beneficial in developing the strategy for natural biological control. It is suggested that the best fortification for the species examined in this study would be a network of water bodies sites, using different traditional and non-intensive methods. This can only be achieved by coordinated planning and conservation measures among sites.

Supplementary material

There is supplementary material associated with this article. Access the material online at: <https://dx.doi.org/10.17582/journal.pjz/20200112020107>

Statement of conflict of interest

The authors have declared no conflict of interest.

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Supplementary Material

Seasonality and Climatic Factors Affect Diversity and Distribution of Arthropods Around Wetlands

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Supplementary Table I. List of arthropods collected over different seasons in two bodies of water in Faisalabad, Pakistan.

Class	Order	Family	Species	Lake				Canal			
				Spring	Sum-mer	Au-tumn	Win-ter	Spring	Sum-mer	Au-tumn	Win-ter
Arachnida	Araneae	Clubionidae	<i>Clubiona pallidula</i>	0	3	0	7	3	9	0	0
			<i>Clubiona spp.</i>	0	3	0	0	0	9	0	0
		Lycosidae	<i>Arctosa spp.</i>	7	5	0	0	12	5	0	0
			<i>Honga lenta</i>	11	7	11	0	6	0	0	0
			<i>Trochosa ruricola</i>	0	0	0	0	9	7	3	8
			<i>Philodromus dispar</i>	3	0	0	0	0	0	0	0
		Philodromidae	<i>Philodromus dispar</i>	3	0	0	0	0	0	0	0
		Anthicidae	<i>Anthicus cervinus</i>	0	8	24	0	0	0	0	0
			<i>Pterostichus anthracinus</i>	4	3	30	0	0	0	0	0
		Carabidae	<i>Pterostichus melanarius</i>	0	0	0	0	0	0	0	18
Insecta	Coleoptera	Chrysomelidae	<i>Aulacophora foveicollis</i>	0	0	0	0	0	0	10	0
			<i>Chrysolina cerealis</i>	0	0	0	0	0	0	12	0
		Coccinellidae	<i>Adalia bipunctata</i>	0	0	0	0	0	0	19	0
			<i>Adalia decempunctata</i>	66	3	0	35	66	6	0	0
			<i>Cheilomenes sexnuculata</i>	0	2	0	36	5	0	0	19
			<i>Coccinella septempunctata</i>	146	0	0	0	72	15	0	17
			<i>Coccinella undecimpunctata</i>	0	0	0	0	15	0	6	0
			<i>Harmonia conformis</i>	0	0	0	0	3	0	0	6
			<i>Micraspis frenata</i>	0	0	0	0	0	0	30	0
			<i>Attagenus brunneus</i>	6	0	0	0	0	0	0	0
		Dermestidae	<i>Attagenus smirnovi</i>	3	9	15	0	0	0	0	0
			<i>Dermestes maculatus</i>	3	15	0	0	0	0	0	0

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Class	Order	Family	Species	Lake				Canal			
				Spring	Sum- mer	Au- tumn	Win- ter	Spring	Sum- mer	Au- tumn	Win- ter
		Elateridae	<i>Elateridae</i> spp.	3	5	0	43	0	0	0	0
		Hydrophilidae	<i>Enochrus blatchleyi</i>	0	0	17	0	0	0	0	0
			<i>Hydrophilus piceus</i>	0	0	29	0	0	0	0	0
		Melyridae	<i>Malachius bipustulatus</i>	0	0	18	0	0	0	0	0
		Ptinidae	<i>Stegobium paniceum</i>	0	40	42	0	0	0	0	0
			<i>Stegobium paniceum</i>	5	31	84	0	0	0	0	0
		Scarabaeidae	<i>Amphimallon majale</i>	0	8	42	0	0	0	0	0
			<i>Aphodius fossor</i>	2	27	0	0	0	0	0	0
			<i>Scarabaeus viettei</i>	0	0	0	0	9	0	0	19
		Staphylinoidea	<i>Dinaraea angustula</i>	6	5	0	12	0	0	0	0
		Tenebrionidae	<i>Crypticus quisquilius</i>	5	9	12	0	0	0	0	0
	Diptera	Agromyzidae	<i>liriomyza flavopleura</i>	2	0	0	0	0	0	0	0
			<i>Adia cinerella</i>	0	0	0	0	0	0	0	6
			<i>Hydrophoria lancifer</i>	5	3	0	0	0	0	0	0
			<i>Anthomyza</i> spp.	9	0	8	0	0	0	0	0
		Asilidae	<i>Asilidae</i> spp.	0	3	0	0	0	0	0	0
			<i>Dioctria oelandica</i>	4	9	0	0	0	0	0	0
		Asteiidae	<i>Asteia amoena</i>	1	0	0	0	0	0	0	0
		Calliphoridae	<i>Calliphora livida</i>	1	3	0	0	0	16	0	0
			<i>Calliphora vomitoria</i>	0	0	0	0	0	0	0	14
			<i>Lucilia sericata</i>	21	6	0	0	119	6	6	6
		Chironomidae	<i>Chironomus plumosus</i>	0	3	15	13	0	0	0	0
			<i>Chironomus</i> spp.	5	3	4	0	0	0	0	0
			<i>Tanytarsus</i> spp.	328	153	153	15	0	13	59	60
		Cicadellidae	<i>Typhlocyba pomaria</i>	12	0	0	0	0	0	0	0
		Culicidae	<i>Anopheles freeborni</i>	0	0	0	0	12	6	6	0
			<i>Anophles labranchiae</i>	2	0	0	0	3	5	12	0
			<i>Culex pilosus</i>	0	0	0	0	35	0	32	34
			<i>Culex pipens</i>	9	6	0	0	1	26	24	0
			<i>Culex tritaeniorhynchus</i>	11	6	0	0	0	0	0	0
		Drosophilidae	<i>Drosophila funebris</i>	0	13	62	0	0	22	43	0
			<i>Drosophila hydei</i>	7	5	63	0	3	23	78	0
			<i>Drosophila melanogaster</i>	3	0	0	12	0	0	0	0
			<i>Drosophila suzukii</i>	6	0	27	0	0	0	0	0
			<i>Sapromyza</i> spp.	33	0	0	0	0	0	0	0
		Muscidae	<i>Musca domestica</i>	35	11	24	16	58	28	96	3

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Class	Order	Family	Species	Lake				Canal			
				Spring	Sum- mer	Au- tumn	Win- ter	Spring	Sum- mer	Au- tumn	Win- ter
			<i>Potamia littoralis</i>	0	3	0	0	0	0	0	0
		Oestridae	<i>Dermatobia hominis</i>	0	0	0	0	0	0	0	5
		Psychodidae	<i>Clogmia albipunctata</i>	0	0	0	0	6	5	0	6
			<i>Psychoda</i> spp.	67	0	0	0	0	0	0	0
		Sarcophagidae	<i>Sarcophaga aurifrons</i>	5	12	0	0	0	0	0	0
			<i>Sarcophaga bercaea</i>	6	0	0	30	1	6	24	0
			<i>Sarcophaga</i> spp.	2	2	0	0	19	0	0	0
		Sciaridae	<i>Bradysia coprophila</i>	0	0	0	0	0	0	0	20
			<i>Bradysia</i> spp.	0	3	0	0	0	0	0	0
			<i>Lycoriella</i> spp.	5	0	0	0	0	0	0	0
			<i>Nemopoda nitidula</i>	0	0	0	0	0	0	12	0
			<i>Sepsis fulgens</i>	12	0	0	0	8	18	0	0
		Simuliidae	<i>Simulium slossonae</i>	0	0	0	0	2	0	0	9
			<i>Simulium</i> spp.	6	0	0	0	0	0	0	0
		Sphaeroceridae	<i>Coproica ferruginata</i>	24	0	0	0	0	0	0	0
		Stratiomyidae	<i>Beris geniculata</i>	0	9	0	0	0	0	0	0
			<i>Pachygaster leachii</i>	4	0	0	9	0	0	0	0
		Syrphidae	<i>Caesar australia</i>	0	0	0	0	0	6	0	0
			<i>Episyrphus balteatus</i>	0	0	0	0	9	18	6	0
			<i>Eupeodes corollae</i>	0	0	0	0	5	0	6	0
			<i>Orthonevra brevicornis</i>	3	0	0	0	0	0	0	0
			<i>Scaeva pyrastr</i>	6	3	0	0	0	0	0	0
		Tachinidae	<i>Pales pavidus</i>	0	3	4	0	0	0	0	0
		Tipulidae	<i>Tipula paludosa</i>	1	3	0	0	0	0	0	0
	Hemiptera	Anamorphaeidae	<i>Symbiotes gibberosus</i>	0	0	0	0	2	0	3	0
		Aphididae	<i>Aphis gossypii</i>	3	0	0	0	0	0	0	0
			<i>Macrosiphum rosae</i>	0	0	0	0	6	0	0	0
			<i>Schizaphis graminum</i>	0	3	0	0	0	0	0	0
		Cicadellidae	<i>Amrasca biguttula</i>	0	0	0	0	73	15	9	0
			<i>Amrasca devastans</i>	0	0	0	0	0	7	0	0
			<i>Cicadulina mbila</i>	0	9	0	0	0	0	0	0
			<i>Evacanthus acuminatus</i>	0	0	0	0	7	6	0	0
			<i>Exitianus exitiosus</i>	4	12	6	0	6	0	0	0
			<i>Exitianus indicus</i>	91	45	96	9	0	0	0	0
			<i>Typhlocyba pomaria</i>	0	0	0	0	18	9	0	0
		Cimicidae	<i>Cimex lectularius</i>	5	0	0	0	0	0	0	0

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Class	Order	Family	Species	Lake				Canal			
				Spring	Sum- mer	Au- tumn	Win- ter	Spring	Sum- mer	Au- tumn	Win- ter
		Erotylidae	<i>Erotylus incomparabilis</i>	0	0	0	0	0	0	0	3
		Lygaeidae	<i>Oxycarenus hyalinipennis</i>	5	3	0	0	0	0	0	0
			<i>Oxycarenus lavatae</i>	1	1	0	0	0	0	0	0
		Miridae	<i>Closterotomus norwegicus</i>	0	0	0	0	0	7	0	0
		Pyrrhocoridae	<i>Dysdercus cingulatus</i>	18	15	0	0	13	11	0	0
			<i>Pyrrhocoris apterus</i>	0	6	4	0	0	0	0	0
	Hymenoptera	Apidae	<i>Apis cerana</i>	0	0	0	27	0	0	0	0
			<i>Apis dorsata</i>	6	0	0	22	7	15	27	21
			<i>Apis florea</i>	8	5	0	12	0	0	0	6
			<i>Apis mellifera</i>	5	6	15	31	48	5	30	2
		Braconidae	<i>Doryctobracon areolatus</i>	0	0	0	0	0	15	0	0
		Formicidae	<i>Camponotus floridanus</i>	46	0	3	53	8	9	0	11
			<i>Camponotus ligniperdus</i>	26	13	3	15	6	0	0	4
			<i>Camponotus pennsylvanicus</i>	9	9	0	6	1	12	3	0
			<i>Camponotus ulcerosus</i>	0	5	0	0	0	0	0	0
			<i>Cardiocondyla</i> spp.s	0	12	0	0	0	0	0	0
			<i>Formica rufa</i>	8	0	0	15	0	0	0	0
			<i>Nomia melanderi</i>	2	0	9	3	0	0	0	0
			<i>Prenolepis imparis</i>	11	0	7	0	0	0	0	0
		Ichneumonidae	<i>Lasius fuliginosus</i>	8	0	0	0	0	0	0	0
			<i>Lasius niger</i>	20	15	15	3	0	0	0	0
			<i>Pimpla</i> spp.	4	0	0	0	0	0	0	0
		Proctotrupidae	<i>Phaneroserphus coreanus</i>	0	3	0	0	0	0	0	0
		Sphecidae	<i>Sceliphron caementarium</i>	0	0	42	36	0	0	0	0
		Vespidae	<i>Polistes carolina</i>	24	11	24	0	0	0	0	0
			<i>Polistes watti</i>	0	20	3	0	5	40	3	0
			<i>Vespa orientalis</i>	6	0	0	0	0	0	0	0
	Lepidoptera	Crambidae	<i>Ostrinia nubilalis</i>	2	0	27	0	0	5	27	0
		Erebidae	<i>Caenurgina erechtea</i>	0	0	0	0	3	0	0	0
		Geometridae	<i>Animomyia</i> spp.	0	3	0	0	0	0	0	0
		Limacodidae	<i>Tortricidia pallida</i>	0	0	0	0	2	0	6	0
		Lycaenidae	<i>Zizula hylax</i>	9	9	0	0	6	6	0	0
			<i>Zizula</i> spp.	3	5	4	0	0	0	0	0
		Noctuidae	<i>Basilodes pepita</i>	0	0	6	0	0	0	0	0
			<i>Oria musculosa</i>	0	0	0	0	0	0	3	2

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Class	Order	Family	Species	Lake				Canal			
				Spring	Sum- mer	Au- tumn	Win- ter	Spring	Sum- mer	Au- tumn	Win- ter
		Nolidae	<i>Garella</i> spp.	0	0	0	0	0	3	0	0
		Nymphalidae	<i>Danaus chrysippus</i>	21	7	15	35	0	0	0	0
		Pieridae	<i>Pieris brassicae</i>	11	9	0	0	6	3	0	0
			<i>Pieris rapae</i>	15	3	0	0	12	3	0	0
		Tineidae	<i>Tineola bisselliella</i>	0	3	0	6	0	0	0	0
		Tortricidae	<i>Epiphyas postvittana</i>	0	0	15	6	0	0	9	6
	Neuroptera	Chrysopidae	<i>Chrysoperla carnea</i>	3	0	0	3	0	0	0	0
		Myrmeleontidae	<i>Myrmeleon formicarius</i>	0	3	0	0	0	0	0	0
	Orthoptera	Acrididae	<i>Acrida conica</i>	6	11	0	3	2	3	0	18
			<i>Acrida exaltata</i>	10	5	8	0	0	0	0	0
			<i>Acrida ungarica</i>	0	0	0	3	3	0	9	3
			<i>Melanoplus bivittatus</i>	3	5	0	0	8	6	10	0
			<i>Melanoplus borealis</i>	0	0	0	0	5	4	0	0
			<i>Melanoplus differentialis</i>	8	0	30	0	9	0	0	7
			<i>Melanoplus femerburum</i>	0	3	0	0	23	3	0	0
			<i>Omocetus viridulus</i>	0	0	0	0	11	0	3	0
			<i>Schistocerca americana</i>	0	0	0	0	2	3	21	6
			<i>Schistocerca lineata</i>	0	3	0	0	0	0	0	0
			<i>Schistocerca nitens</i>	2	0	0	3	5	5	6	9
			<i>Schistocerca</i> spp.	15	19	0	0	0	0	0	0
			<i>Acheta domesticus</i>	0	11	0	0	0	0	0	0
			<i>Gryllus campestris</i>	0	0	0	0	0	3	0	0
		Lygaeidae	<i>Omocetus viridulus</i>	11	0	7	0	0	0	0	0
		Pyrgomorphidae	<i>Poecilocus pictus</i>	0	3	0	0	0	0	0	0
		Rhaphidophoridae	<i>Diastrammena asynamora</i>	0	0	0	0	0	0	3	1
		Tettigoniidae	<i>Scudderella mexicana</i>	0	0	0	0	7	6	6	2